

**COMPARISON OF POPULATION EXPOSURES DURING FOUR SELECTED  
EPISODE DAYS IN HELSINKI IN 2002**

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**INTRODUCTION**

Ambient urban air pollution, especially fine particulate matter (PM<sub>2.5</sub>), has been associated to excess mortality and morbidity at the current urban levels (WHO, 2000). Urban air quality is regulated for protection of public health using limit values to which prevailing air quality is compared. Urban populations, however, are mobile and spend their time in different parts of their cities throughout the day, including large fractions of time spent in indoors, where the building modifies air quality (Koistinen *et al.*, 2001). A smaller but equally important fraction of time is spent in traffic, where the actual exposures are typically much higher than elsewhere. Therefore in order to evaluate the adverse health effects of air pollution, urban air quality models need to be complemented with modelling of population exposures.

In the current study the FMI urban air quality modelling system for Helsinki was linked with a previously developed probabilistic exposure model. The modelling system was then used to simulate working age population exposure distributions for four selected episode days: spring dust episode (10.4.2002), long-range transport episode (14.4.2002), long-range transport episode with a special contribution for vegetation fires (28.8.2002), and local inversion (22.10.2002). The objectives of the current work are A) evaluation of dispersion and exposure model performance for episodes, B) compare exposure levels during different types of episodes, C) study the relationship of monitored ambient air quality and population exposures, and D) evaluate the health relevance of population exposures in episode situations.

**METHODOLOGY**

A probabilistic model was previously developed for simulation of population exposures to fine particles (Hänninen *et al.*, 2003; Kruize *et al.*, 2003; Hänninen *et al.*, 2005). The model is based on microenvironment approach. Working age population time activity was classified into five microenvironments for the simulation (residence, workplace (working sub population), other indoor environments, outdoors, and in-traffic (including walking and cycling)). Population time-fraction spent and concentrations in each microenvironment were input as parameters of beta and lognormal probability distributions, respectively.

Dispersion model computations (Karppinen *et al.*, 2000a, b) were used to estimate hourly spatial distributions of PM<sub>2.5</sub> outdoor concentrations during the selected episode days. A stratified random sample of residence and workplace locations was drawn to model spatial variability in population densities between the city areas for different age groups. Lognormal distributions were fitted to the data and used in the simulation.

The indoor PM<sub>2.5</sub> concentration of ambient origin was calculated using the infiltration factors (Hänninen *et al.*, 2004) and outdoor concentrations as described in more detail in Hänninen *et al.*, 2005. Indoor concentrations were included in a model validation run to compare with corresponding measurements (Fig. 1-A). Because the focus of the present work is on ambient air pollution episodes, all exposure levels refer to PM<sub>2.5</sub> of ambient origin, and in the final models indoor sources were set to zero. Concentrations experienced while in traffic were estimated using measurement data and a state-of-art data analysis (Jantunen *et al.*, 2005).

The dispersion model predictions were compared with monitoring station data. For the whole year of 2002 the daily averages agreed fairly well, but for the episodes the model underestimated levels at fixed monitoring sites in all four cases. For estimation of true exposures, the dispersion model results were corrected by using the fixed monitoring data from Vallila site (Table 1). In case of episodes of local origin (spring dust and inversion) it was assumed that the underestimation of concentrations is relative. For long-range episodes it was assumed that adding the Vallila difference corrects all levels.

**Table 11. Comparison of predicted and observed levels at Vallila station. Terms used for correction of spatial model data are shown in bold.**

Vallila		a-spring dust 10.4.2002	b-LRT 14.4.2002	c-LRT/Fire 28.8.2002	d-inversion 22.10.2002
Observed level	µg m <sup>-3</sup>	27.1	32.0	41.5	24.7
Predicted level	µg m <sup>-3</sup>	24.6	20.6	20.3	13.0
Ratio obs/pred	1	<b>1.10</b>	1.55	2.04	<b>1.90</b>
Difference obs-pred	µg m <sup>-3</sup>	2.5	<b>11.4</b>	<b>21.2</b>	11.7

The exposure model was compared against personal exposures measured in 1996-97 in the EXPOLIS Helsinki -study (Jantunen *et al.*, 1998; Koistinen *et al.*, 2001) for quantification of the model errors. The simulations were executed using the EXPOLIS simulation framework (Hänninen *et al.*, 2003; Kruize *et al.*, 2003) with Microsoft (Seattle, WA) Excel version 8.0 and @Risk add-on software version 4.0 (Palisade, Newfield, NY). The selected simulation settings included latin hypercube sampling, 500 iterations and a fixed pseudo-random number seed. Lognormal distribution fits were truncated at 99.9th percentile.

## RESULTS AND DISCUSSION

**Model performance.** Dispersion model underestimated the levels during all episodes (Table 1; preliminary evaluation see Hänninen *et al.*, 2004b). Therefore additional exposure models were run with corrected input values for estimation of true exposure levels. The exposure model itself was evaluated against previous personal exposure measurements, including estimated indoor sources in the model. The overall match with the observations was good. Model errors (differences between the solid area and line in Fig. 1-A) caused by the simplifications in the model were estimated as the relative difference between the bias-corrected observed percentile and the corresponding simulated value. Between the 5th and 95th percentiles the relative model errors varied between -2 and +20 % and absolute errors between -0.3 and 1.3 µgm<sup>-3</sup>. In the tails of the distribution both the absolute as well as relative errors increase. Especially in the high-end tail the exposure levels are clearly underestimated, indicating relatively rare high personal exposures that are captured with this simple five-microenvironment model. Reasons for underestimated high-end exposures

include missing high concentration microenvironments and time spent in proximity of sources, where the general concentration in the current microenvironment is not representative of the exposure level. Some of the underestimated high exposures might be attributable to misclassified exposures to tobacco smoke.

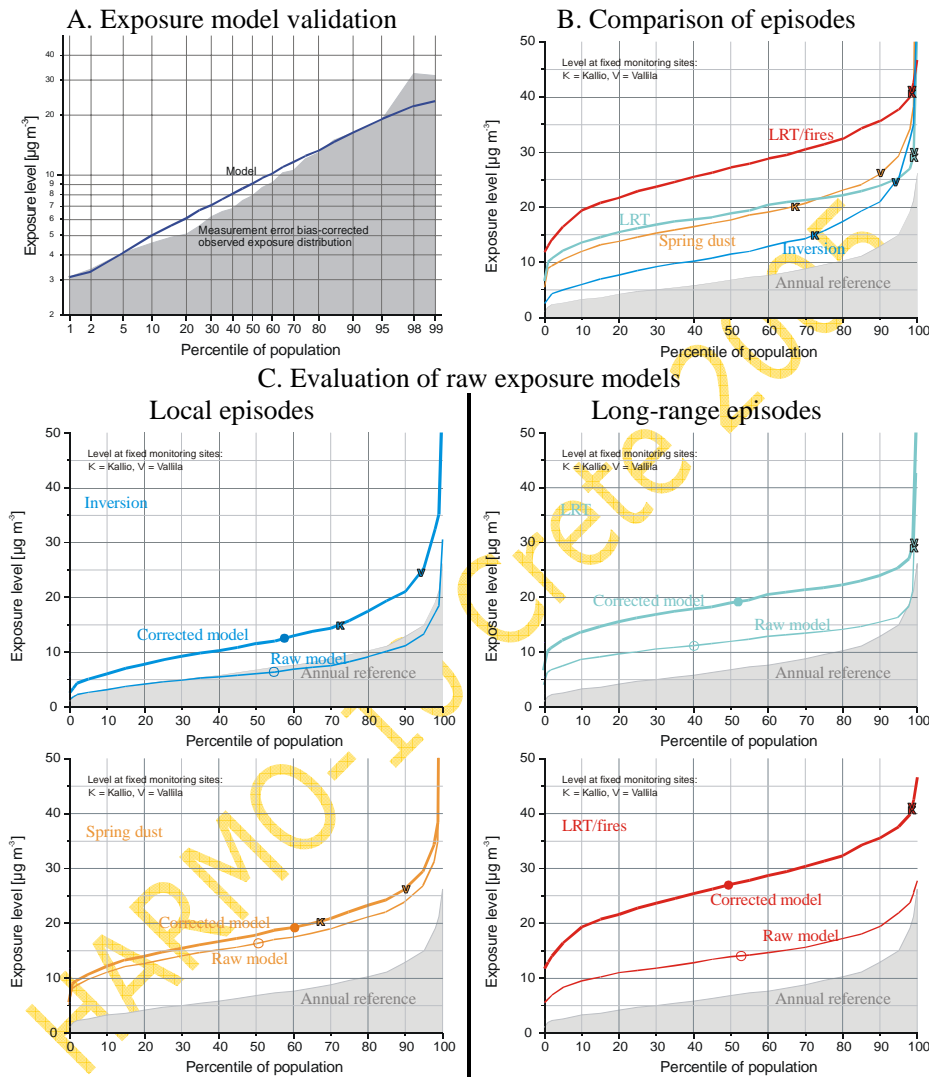


Fig. 23. Exposure results; A. Validation of the exposure model against exposure measurements; B. Comparison of the corrected exposures for the different episodes; C. comparison of raw and corrected models for each episode.

**Comparison of exposures during different episodes.** Highest exposures occurred during the vegetation fire episode and lowest during the inversion day (Fig. 1-B, Table 2). The mean exposure level of the former ( $27 \mu\text{g m}^{-3}$ ) is double compared to the latter ( $13 \mu\text{g m}^{-3}$ ). The spring episodes from local dust and long-range transport are very close to each other. The local inversion episode captured in this study does not represent a worst case, but worse inversions are rare. In contrast, long-range episodes occurred during the study year several

times and are known to be common. In this sense the current results emphasize the role of long-range transported particles in formation of population exposures.

**Table 12. Evaluation of the simulation model performance and mean exposure levels.**

Exposures and concentrations	E p i s o d e							
	Spring dust		LRT		LRT/fires		Local inversion	
	10.4.2002		14.4.2002		28.8.2002		22.10.2002	
	mean ± sd	mean	sd	mean	sd	mean	sd	
	[µg m <sup>-3</sup> ]	[µg m <sup>-3</sup> ]	[µg m <sup>-3</sup> ]	[µg m <sup>-3</sup> ]	[µg m <sup>-3</sup> ]	[µg m <sup>-3</sup> ]	[µg m <sup>-3</sup> ]	
<b>SIMULATION</b>								
Raw dispersion model based exposure	17.1 ± 6.4	12.0 ± 3.2		14.3 ± 4.0		6.8 ± 3.6		
Corrected exposure <sup>1</sup>	18.8 ± 7.0	19.0 ± 4.4		27.2 ± 6.4		13.0 ± 6.8		
Raw model exposure error	-1.7	-0.6	-6.9	-1.2	-12.9	-2.4	-6.1	
Relative error %	-9 %	-9 %	-37 %	-28 %	-48 %	-37 %	-47 %	

**Monitored air quality compared with exposures.** Comparing daily average concentrations observed at the two available fixed monitoring sites (Kallio and Vallila; symbols K and V in Fig. 1-B and C), it is evident that during the LRT episodes both monitored ambient levels represent high exposure percentiles. On the other hand, for the episodes of local origin, spring dust and inversion, the Kallio station level is close to 67<sup>th</sup> and 73<sup>rd</sup> percentiles and the Vallila level close to 90<sup>th</sup> and 95<sup>th</sup> percentiles, respectively. The Kallio level is quite close to the mean (108% and 113% for the spring dust and inversion episodes, respectively). During the LRT episodes the fixed station levels are close to each other, as can be expected, and both overestimate the mean exposure level approximately 1.5 fold. Moderately traffic oriented Vallila station overestimates mean population exposures during all episodes.

**Exposure levels and their health relevance.** Exposure distributions for all episodes were significantly higher than the annual reference distribution (1996-7 exposure from the validation model without indoor sources; Fig. 1-B). Mean exposure level during the vegetation fire originating long-range transportation episode was three times higher than the annual reference. According to the current understanding, the health effects are caused primarily by accumulating long-term exposures and only secondarily by short-term (i.e daily) exposures; the exposure-response factor for PM<sub>2.5</sub> for the former is about an order of magnitude higher than for the latter (WHO, 2000). Short-term exposures are, however, very relevant for the currently developing EU legislation requiring regulatory actions in case of episodes. Exposure and health research is needed for optimized use of resources and maximal health benefits.

## CONCLUSIONS

Environmental policies should ultimately target at improved public health while minimizing non-productive intervention and cost. This can be achieved by optimization of air quality management decisions according to the quantitative estimates of exposures and health risks in scenarios for alternative policy options. Using exposure models in an episode situation makes it possible to compare exposure reduction schemes in terms of efficiency and other local considerations for optimal decisions, including limitations for e.g. private car traffic or local industrial activities.

The same models are valuable in long-term scale. Population exposures can be estimated for e.g. alternative traffic system scenarios, as is done in Helsinki as part of the metropolitan area transportation system plan. The modelling tools can be used to apportion exposures to

sources, including the long-range transported pollution, and thus to support setting priorities for local, regional, and international emission control activities.

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